Biology and Impacts of Pacific Island Invasive Species. 2. Boiga irregularis, the Brown Tree Snake (Reptilia: Colubridae)

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Biology and Impacts of Pacific Island Invasive Species. 2. *Boiga irregularis*, the Brown Tree Snake (Reptilia: Colubridae)

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**Abstract:** The Brown Tree Snake, *Boiga irregularis* (Merrem, 1802), was accidentally transported to the island of Guam shortly after World War II. Over the following two decades it spread throughout the island with little public or professional recognition of its extent or impacts. This secretive nocturnal arboreal snake occurs in all habitats on Guam, from grasslands to forests. Under the right conditions, it is capable of high rates of reproduction and population growth. The Brown Tree Snake caused the extirpation of 13 of Guam’s 22 native breeding birds and contributed to the extirpation of several species of native bats and lizards. Guam’s 12 forest birds were especially impacted, with 10 species eliminated and the other two severely reduced. In addition, the snake continues to substantially impact domestic poultry, pets, the island’s electrical power infrastructure, and human health. To protect other vulnerable Pacific islands, the U.S. government annually spends several million dollars inspecting cargo outbound from Guam to exclude Brown Tree Snakes. Cargo destinations most at risk are in Micronesia, especially the Northern Mariana Islands, but Guam also has direct air transportation links to Hawai’i that will soon be supplemented with direct ship traffic. Ultimately, all Pacific islands are at risk but especially those obtaining cargo through Guam.

The Brown Tree Snake, *Boiga irregularis* (Merrem, 1802), is not an especially unusual snake, though it is longer (up to 3 m total length), skinnier, more nocturnal, and more arboreal than an average snake. Its notoriety derives from the profound ecological impacts it had on vertebrate life upon its postwar arrival on the island of Guam. In this paper we review the history of the snake’s impacts on Guam and outline management efforts to prevent new invasions by this species on other Pacific islands and, if possible, to reverse the ecological and economic dislocations witnessed on Guam.

**Name**

*Boiga irregularis* is most often called the Brown Tree Snake, though a variety of names are in local use (Rodda et al. 1999b). Our favorite is northern Australia’s “Doll’s Eye,” which recognizes the disproportionately large eyes of this species, especially when it is young (Kinghorn 1964).

The Brown Tree Snake is one of about 1,800 species of modern snakes lumped into the poorly resolved and presumably polyphyletic family of “typical” snakes: Colubridae. A cluster of similar genera is sometimes characterized as the subfamily Colubrinae (formerly placed in Boiginae). The genus *Boiga* is an Asian radiation of slender rear-fanged snakes, with 33 described species (EMBL [European Molecular Biology Laboratory] Web site accessed 31 January 2006: http://www.embl-heidelberg.de/~uetz/LivingReptiles.html); in-
sofar as it is known, all members of the genus are nocturnal and all but one are arboreal. The easternmost representative of this Asian radiation is the Brown Tree Snake. As currently recognized, *Boiga irregularis* includes all of the closely related brown *Boiga* found east and south from Sulawesi to Australia and the Solomon Islands (see a later section on Geographic Distribution), though various proposals have been suggested to recognize locally distinctive forms. The species name *irregularis* recognizes the diversity of morphology in the species.

**DESCRIPTION AND ACCOUNT OF VARIATION**

**Species Description**

The Brown Tree Snake conveys the impression of having a large head in relation to the diameter of its neck and body. For a snake of this length, the head size is actually normal, but the body is extraordinarily slender (Rodda et al. 1999b). Any hole that will admit a Brown Tree Snake’s head will allow passage of the entire snake (unless a very large meal has recently been ingested). The snout is short and the eyes are large, with the elliptical pupil characteristic of nocturnal species (Figure 1). The teeth are of moderate length, the upper rearmost of which are enlarged and grooved to facilitate penetration of venom into prey (Vest et al. 1991, Hayes et al. 1992, 1993, Jackson and Fritts 1995, Mackessy et al. 2006). The venom is relatively mild for humans (Weinstein et al. 1991, 1993), though it overlaps with the toxicity of that of species considered dangerous (e.g., North American Copperheads, *Agkistrodon contortrix*). Compared with the copperhead, however, the Brown Tree Snake lacks a means for injecting the venom (only capillary action conveys the venom into prey). Toxicity varies greatly among prey species, with select birds and lizards being highly susceptible to neurotoxic elements of the venom (Mackessy et al. 2006).

**Figure 1.** A juvenile Brown Tree Snake from Guam, showing the Guam population’s characteristic shade (light/medium brown) and vague black banding, as well as the species’ characteristic defensive pose, short snout, and large eyes. (Photo by G. Rodda)
Brown Tree Snakes have tails that average 21% of their total length. This proportion is intermediate among snakes and relatively short compared with most arboreal snakes (Guyer and Donnelly 1990). No sexual size dimorphism in relative tail length or width has been reported, though male Brown Tree Snakes often achieve a larger total length than do female Brown Tree Snakes (Savidge 1991, Rodda et al. 1999b). The degree of sexual size dimorphism is sensitive to the sampling method. The largest recorded male on Guam was 3.1 m in total length, whereas the largest recorded female was only 1.9 m. These lengths are exceptional, however, the 99th percentile sizes on Guam are 2.1 m ($n = 4,870$) and 1.5 m ($n = 5,223$) for males and females, respectively. These sizes are larger than those reported for most of the native range, especially Australia, where maximum total lengths are typically reported to be 1.8 m (males). On Guam, the smallest hatchlings that have been found are about 330 mm snout-vent length (SVL), whereas Australian hatchlings are reported to be smaller (250–275 mm SVL [Gow 1976, Shine 1991, Greer 1997]).

In coloration and scutellation, *Boiga irregularis* rangewide lives up to its “irregular” description, while having relatively invariant coloration and scutellation at any one locality (Figure 2). Typically the species has vague to distinct narrow to wide blackish banding dorsally on a medium brown background, countershaded to a cream to yellowish venter flecked with tiny lighter and darker marks. However, in parts of Australia the banding may be whitish, blue, or red. Details of scut-

![Figure 2](https://bioone.org/journals/Pacific-Science/309/PacificScience309_f2.jpg)
Distinguishing Features

A combination of traits rather than any single diagnostic character should be considered:

1. Size: SVL 250–2,700 mm
2. Color pattern: banded or unmarked; brown, blue, or reddish brown background color
3. Head shape: blunt short snout with wide quadrates (relative to neck) and large eyes
4. Body shape: very slender, with mass about 100 g for a 1,000-mm SVL individual
5. Head scation: typical of colubrids
6. Body scation: often but not always with transversely enlarged middorsal scales
7. Dorsal scale rows in midbody region: 17–25
8. Anal scale: undivided or divided
9. Number of ventrals: 217–286
10. Number of subcaudals: 65–130

Intraspecific Variation

As already noted, the coloration and scutellation of the Brown Tree Snake is variable from place to place. Only in Australia have systematists attempted to split the species or recognize subspecies, treating forms in the northwest of Australia as subspecies of the full species _fusca_ (Kinghorn 1964, Storr et al. 1986, Ehmann 1992, Whittier et al. 2000). The distinguishing characteristics are inconsistent from observer to observer, however, and different geographic boundaries of division have been proposed (Rodda et al. 1999b). Furthermore, the defining morphological characters have not been mirrored in molecular evidence of gene flow restriction, suggesting a lack of genetic structuring in Australia (S. Donnellan and others, pers. comm.). Most island isolates of Brown Tree Snakes do show genetic structure, but no distinctive forms have been formally recognized.

Detrimental

Economic impacts from the Brown Tree Snake on Guam include damage to electrical power infrastructure, loss of pet and domestic animals, human envenomations, higher costs of shipping from Guam, and threats to the tourism industry.

The most quantified economic factor has been electrical system costs (Fritts et al. 1987, Fritts 1994, 2002, Fritts and Chiszar 1999, Burnett et al. 2004). Burnett et al. (2004) extrapolated from the electrical system damage experienced on Guam to the expense associated with a similar level of outages throughout the state of Hawai‘i, estimating a staggering $1.7 billion in annual costs. Based on the recovery of electrocuted snakes at the site of electrical “faults,” Guam experiences a snake-caused power outage about every other day on average, but many of the outages affect only a small area. Costs include direct damage to the electrical infrastructure (damaged transformers, generators, etc.), emergency restoration costs, loss of revenue to the power company while power is not being delivered, lost customer productivity while power-dependent systems (e.g., computer networks) are down, disruption of normal urban functions (e.g., elevator outages, traffic jams associated with nonfunctioning traffic signals), and the costs associated with mitigation (requirement for extra generators, etc.). No analysis has considered all of these costs simultaneously, and many are dependent on poorly documented assumptions about how electrical power systems would work in the absence of Brown Tree Snakes. Nonetheless, it is clear that the economic costs of frequent power outages are a substantial burden on business and residential customers.

Brown Tree Snakes consume many pet animals on Guam, but this impact has not been quantified. Anecdotally, the loss is primarily to neonatal mammals, especially puppies, and cage birds. The primary loss of domestic animals is of poultry (Savidge 1987a, Fritts and McCoid 1991). Loss of local suppliers of eggs requires costly importation by air. It is
difficult to assign exact costs, because this requires making assumptions about the state of Guam’s poultry industry in the absence of snakes.

The primary victims of envenomations are infants sleeping at home in their cribs (Fritts et al. 1990, 1994). A similar pattern of infant envenomations is recorded for Asian residents of areas inhabited by the elapid snakes of the genus *Bungarus* (Fritts and McCoid 1999). Adult humans appear not to be very vulnerable to Brown Tree Snake venom, because they remove the biting snake before capillary action introduces substantial venom. Although no human fatalities have been recorded for bites by *Boiga*, at least 10 Guam infants have been placed on ventilators or otherwise put on life support while being treated. The economic cost of such treatment is presumably minor compared with the psychological trauma associated with possible loss of a child.

These economic burdens increase the cost of attracting desired employees and conducting business on Guam. Inspections of cargo leaving Guam undoubtedly boost the cost of shipping, but most of the cost is now borne by the mainland taxpayer. Guam shippers voluntarily arrange and wait for inspections by U.S. Department of Agriculture Wildlife Services, but this burden is certainly buried in shipping fees. Various plans have been suggested to assess shippers a fee for the inspection costs now carried by the federal taxpayer, and these would further increase the expense of shipping from Guam.

The military community experiences many of the same burdens from the Brown Tree Snake as do civilians (power outages, envenomations, loss of pets, difficulties recruiting employees from off island, etc.), and it also pays Wildlife Services for the costs ($1—2 million/yr) of inspecting materials shipped from Guam. The U.S. Department of Defense is the source of the majority of cargo leaving Guam (Daniel Vice, pers. comm.).

Guam’s economy is primarily dependent on tourism from Asia (Guam Department of Commerce 1998). Although few Guam tourists personally encounter Brown Tree Snakes, ophiophobia is based more on fears and impression than coldly calculated risks. No quantification has been published for lost tourist revenues. Tourists can easily change their destination from Guam to snake-free sites such as Saipan or Hawai’i that still host a variety of birds. Preliminary cost estimates for tourism impacts should the Brown Tree Snake reach Hawai’i are in the range $0.5—1.5 billion annually (S. Shwiff, pers. comm.).

Direct ecological costs of the introduction of the Brown Tree Snake to Guam have been relatively well documented, but indirect losses (e.g., loss of forest trees due to absence of essential seed dispersers) have been largely overlooked. The direct losses include extirpation of most birds (Savidge 1987b, Rodda et al. 1997, Fritts and Rodda 1998), population reduction of the flying fox (Wiles 1987, 1989), and extirpation or reduction of several lizard species (Rodda and Fritts 1992a). The cumulative effect of these losses is quite striking; there are areas on Guam that lack all native vertebrates except for a few species of small lizards. Thus, processes that involve vertebrates, such as insectivory, frugivory, seed dispersal, pollination, and processes dependent on these phenomena, have undoubtedly been disrupted, the consequences of which have not yet been explored. However, quantification of these disruptions is complex, in part due to introductions of other species, such as ungulates and various invertebrates, and the impacts of numerous typhoons on vegetative structure. Typhoons are natural, but climate change may have elevated the frequency of typhoons in recent decades (U.S. Fish and Wildlife Service 2005).

**Beneficial**

Introduced small mammals (including rats) are less numerous than they were before the arrival of the Brown Tree Snake (Savidge 1987b).

**Regulatory Aspects**

To date, Brown Tree Snake regulation has been mostly voluntary, with only a requirement for a permit needed from the U.S. Fish and Wildlife Service for importation for scientific or exhibition purposes. Importation
for pet ownership is not permitted in the United States. However, the primary risk of spread is not through intentional transport but through accidental movement. No regulatory requirements exist regarding cargo sanitation or inspections, though cargo inspection certification has been much discussed among Guam and its trading partners. The U.S. Congress requested a study of possible cargo inspection certification (7 USC 8504) through the Brown Tree Snake Control and Eradication Act of 2004 (P.L. 108-384), but participation in any such program is currently voluntary.

**GEOGRAPHIC DISTRIBUTION**

The distribution of the Brown Tree Snake includes its native range of coastal Australia, Papua New Guinea and a large number of islands in northwestern Melanesia (Fritts 1988), and the introduced population on Guam (Table 1). The Brown Tree Snake is found on almost every island from Sulawesi to Guadalcanal, and south across the Torres Strait to the northern and eastern coasts of Australia (see Rodda et al. 1999b for a range map). It is found on a few outlier islands as well (e.g., Santa Cruz Islands in the Solomons), where it may have been prehistorically introduced by humans. However, the only unequivocally established introduced population is on Guam. Persistent reports from the island of Saipan (Fritts et al. 1999) suggest that the snake may be in the process of becoming established there, but clear evidence of establishment or recruitment is lacking. Sightings on Saipan in the last 5 yr have clustered just west of the airport, though earlier specimens were collected at several sites away from the port or airport (N. Hawley, pers. comm.). All but one of eight specimens from Hawai‘i predate the 1994 initiation of Wildlife Service’s control program on Guam (Fritts et al. 1999), and the Hawai‘i sightings have not been clustered in time or space. An old record for Java (Capocaccia 1977) needs corroboration; if documented this would be a second extralimital population. Published records or museum specimens are known for the islands listed in Table 2.

**HABITAT**

**Climatic Requirements and Limitations**

The habitat requirements or habitat-selection criteria of this species have not been studied, although the snake’s population biology and movements have been quantified in several habitats. Its distribution suggests difficulty in surviving in sites exposed to hard frosts; the southern limit of Brown Tree Snakes lies in the suburbs of Sydney, Australia, at the northern limit of hard frosts. In New Guinea the upper-elevation limit is the altitudinal frost line (Greer 1997). Brown Tree Snakes hibernate in the higher latitudes of their range (Covacevich and Limpus 1973, Hoser 1980), which suggests that it is not direct exposure to low temperatures that is limiting. Low prey availability or other factors (vulnerability to predation when cool) represent potential benefits of inactivity during winter.

Brown Tree Snakes do not occur in the arid interior of Australia, even where the prevailing temperatures are within the range tolerated by the snake in more coastal areas. The reason for this limitation is not understood. Relative unavailability of prey, a shortage of trees, or simple desiccation may be responsible. There is some evidence for each of these. To the best of our knowledge, data to contrast prey availability between coastal and interior Australia are unavailable. Morton and James (1988) quantified lizard prey availability in the arid interior (2.9 kg/ha), which is an order of magnitude lower than lizard prey availability in the Marians (mean = 30.0 kg/ha, range 7–65 kg/ha [unpubl. data]), but comparable figures are not available for coastal areas of Australia. Brown Tree Snakes are associated with trees, which are scarce in central Australia. The species utilizes savanna habitat on Guam, where desiccation is presumably reduced due to shading in the tall, dense grass. Furthermore, Guam’s grasslands are interdigitated with strings of trees. Anderson (2002) and Anderson et al. (2003) found that Brown Tree Snakes from Guam are relatively vulnerable to desiccation. However, no Brown Tree Snakes from arid areas (such as Australia) have been tested. Thus many factors could impinge on the abil-
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**Habitat and Resource Requirements and Limitations**

On Guam, where climate, predation, and prey are not limiting, Brown Tree Snakes occupy all habitats (native forest, introduced forest, agroforest, shrubland, rocklands, grassland, suburban, and urban). Habitat, independent of prey availability, does not seem to be a key factor in their distribution. However, we do not fully understand their requirements through all phases of their life cycle, especially breeding; they may need to move among key microenvironments (Fritts and Rodda 1998). Juvenile Brown Tree Snakes...
eat exclusively lizards (Savidge 1988, Greene 1989, Shine 1991), and lizards are more available in some habitats than others, but these differ from place to place. Similarly, adult Brown Tree Snakes on Guam are now heavily dependent on introduced rodents, which often have different habitat requirements than lizards.

Predation is not an important limiting factor for Guam populations of the Brown Tree Snake. Although feral domestic predators (cats, dogs, and pigs) take Brown Tree Snakes opportunistically, almost all recorded snake mortality is associated with starvation and low individual body weight (G.H.R., K. Dean-Bradley, T. H. Fritts, and J.A.S., unpubl. data). Average body condition of snakes throughout Guam declined irreversibly in all habitats coincident with the peak of the Brown Tree Snake irruption (early 1980s), though snakes in grassland habitats exhibit consistently better body condition than those in other habitats. What predation there is from feral animals on Guam is not obviously correlated with any particular habitat.

Prey availability is important to the success of Brown Tree Snakes. An analysis of Brown Tree Snake densities in relation to various local attributes indicated that all of the explained variation (83%) was attributed to availability of different types of prey (Rodda et al. 1999c). Many small islands of the world, including the major islands in the Marianas, show exceptionally high prey densities (Rodda and Dean-Bradley 2002); if this trend holds uniformly, we expect that Brown Tree Snakes could thrive on most tropical or subtropical islands.

### Ecosystem and Community Types Invaded

To date (2007) Brown Tree Snakes have been documented to colonize only one locality: Guam (though Java is an undocumented second possibility). Because the snake spread rapidly throughout the available land area on Guam, and no failed colonizations are known, we have no data on any limitations associated with ecosystem or community types although population densities vary with habitat type.

### History

Brown Tree Snakes arrived on Guam around 1949, most likely as a passive stowaway on World War II war materiel being salvaged from the New Guinea area through the large port facility at Manus Island (Savidge 1987b, Rodda et al. 1992). Through interviews and historical documents, Savidge (1987b) docu-
mented the spread across Guam. The first snakes were noticed near the Fena Reservoir/Santa Rita area of Guam. They spread concentrically from that location, reaching the farthest point of the island (Ritidian Point, ca. 37 km from the point of initial colonization). The date of arrival in the north is not well documented; one record was obtained in 1968, but the next records were from 1982 (Savidge 1987b). Presumably populations were at low densities before general detection by the public. Brown Tree Snakes are usually arboreal, nocturnal, and cryptic; thus their presence could be easily overlooked.

Extrapolating from capture rates achieved during the 1980s (Rodda et al. 1992), and recognizing that small snakes were not caught by the traps then (or now) in use (Boyarski 2005, Rodda et al. 2007), the peak density of Brown Tree Snakes in favorable habitats was probably in excess of 100/ha. Such a density is unprecedented for a nonaggregated large snake (Rodda et al. 1999c), which average around 2/ha (range 0.1–14 [Parker and Plummer 1987]).

Snakes throughout Guam grew abruptly skinnier in the early 1980s, suggesting that they had exceeded their carrying capacity nearly simultaneously islandwide (G.H.R., K. Dean-Bradley, T. H. Fritts, and J.A.S., unpubl. data). Coincident with this decline in body condition was the extirpation of most native prey species (Savidge 1987a, 1991). What permitted the snake to thrive on Guam following the loss of most native prey species? Continued survival was possible because the snake’s abundance was supported by abundant introduced prey species, especially a skink (Carlia ailanpalai), geckos (especially Gebyra sp.), rodents (Rattus spp. and Mus musculus), a shrew (Suncus murinus), and a variety of introduced birds (Columba livia, Streptopelia bitorquata, and Paser montanus) (Rodda et al. 1997, Fritts and Rodda 1998).

By the mid 1990s, the snake’s populations on Guam appeared to have reached a dynamic equilibrium, at roughly half the estimated peak density in the 1980s (Rodda et al. 1992 and unpubl. data). No directional change in density has been observed since that time.

**Physiology**

Ecophysiology of snakes tends to focus on thermoregulation, with subsidiary interests in water balance, metabolism, and sensory systems, especially sensory systems crucial for food acquisition. In all respects the physiology of the Brown Tree Snake is nonremarkable, though not necessarily representative of a typical snake. For example, the Brown Tree Snake is a thermal conformer. Anderson and colleagues studied thermoregulation in the Brown Tree Snake by radiotelemetry in the wild and in thermal gradients. A wide variety of temperatures was experienced in both situations (Anderson 2002, Anderson et al. 2005). Brown Tree Snake skin has relatively high permeability to water, suggesting a vulnerability to desiccation. To the extent it is known, Brown Tree Snake metabolism appears unremarkable.

As suggested by the conspicuous eyes of the Brown Tree Snake, this species is perhaps more dependent on eyesight than many snakes. The balance between dependence on chemical and visual cues has been the subject of a long series of experiments by Chiszar and colleagues (Chiszar et al. 1985, 1988a, b, 1999, Chiszar and Kandler 1986, Chiszar 1990, 1992, 1999; see also Gee 2002). The most novel finding is that Brown Tree Snakes will cease pursuit of a prey item that can be seen to be absent, whereas some, perhaps many, other snakes will continue pursuit if chemical cues remain (Weldon et al. 1994, Schwenk 1995). Studies of relative importance of various cues have been impeded by chronic discrepancies between the behavior of Brown Tree Snakes in captivity and in the wild (Chiszar et al. 1997). Despite the propensity of wild Brown Tree Snakes to thrive in close proximity to human habitation, Brown Tree Snakes appear to be unusually stressed by captivity (Mathies et al. 2001, Aldridge and Arackal 2005, Moore et al. 2005).

**Reproduction**

Probably no aspect of Brown Tree Snake biology has been more frustrating to scientists than reproduction. Although this species
is obviously a successful reproducer, it has proven extremely difficult to breed under controlled conditions. The most likely explanation is that the species’ reproduction is exceptionally vulnerable to human disturbance (Mathies et al. 2001, Aldridge and Arackal 2005, Moore et al. 2005). A few attempts at captive propagation have been partially successful (Barnett 1993, Greene and Mason 2000, Mathies and Miller 2003). More often courtship has been observed without it necessarily progressing to viable eggs (Greene and Mason 1998, Greene 1999, Mathies et al. 2004). These and other studies support the impression that chemical cues play a central role in courtship (Greene et al. 2001, 2002, Mason and Greene 2001, Greene and Mason 2003, 2005), though success at isolating the responsible chemicals has not been reported (Murata et al. 1991), with the negative evidence implying that a bouquet of chemicals rather than a single compound is involved.

The most novel finding to emerge from this body of work is that Brown Tree Snakes can be induced to breed by cooling to temperatures not found on Guam (Greene 1999, Greene and Mason 2000, Mathies and Miller 2003, Mathies et al. 2004). This is consistent with Hoser's (1999) report of winter breeding at the extreme southern limit of the species’ range.

Breeding on Guam takes place year-round (McCoid 1994, Savidge et al. 2007), with some evidence for peak hatching in the wet season (Rodda et al. 1999b); thus peak copulation and oviposition probably occur in the preceding dry season. However, some recruitment occurs in every month of the year on Guam, at least in some years (Savidge et al. 2007). Guam’s climate is moderately seasonal, with least rain in March (78 mm) and most in October (411 mm); annual temperature variation is only 1.4°C (Savidge et al. 2007). Year-round reproduction is an impediment to contraceptive control of the Brown Tree Snake, because any temporary inhibitors would need to be administered continuously. In the more temperate or seasonal parts of the snake’s range, reproduction is seasonal (Shine 1991, Whittier and Limpus 1996, Bull et al. 1997).

Population dynamics

We can estimate a potential rate of spread or population expansion for Brown Tree Snakes that can be used cautiously to anticipate control needs for other localities at risk. As indicated in the section on History, the initial spread of Brown Tree Snakes on Guam covered 37 km (the distance from the point of original infestation to Ritidian Point) in potentially less than 20 yr. We do not know that there was more than one locus of infestation, as might occur if human transport facili-
tated the spread of the snake across Guam, but neither can that possibility be entirely ruled out (either on Guam or for prospective infestations). Using 19 yr as a possible minimum, the estimated linear rate of spread would be 1.95 km/yr or about 2 km/yr. Peak density was not achieved at Ritidian Point until shortly after 1982. If that scenario were to hold for a larger island (e.g., O’ahu), it might take many decades for a Brown Tree Snake irruption to peak, whereas for a smaller island (e.g., Rota, just north of Guam), the peak irruption might be achieved much sooner.

We can roughly estimate the population growth of the Brown Tree Snake population on Guam with reference to two landmarks: first colonization and the estimated population in the early 1980s. Based on finding no variability in mitochondrial DNA, L. H. Rawlings (1995 and pers. comm.) declared that all Brown Tree Snakes then on Guam descended from a single female colonist. That female was probably on Guam around 1949. There were an estimated 2 million Brown Tree Snakes on Guam in the early 1980s (population estimation methods given in Fritts and Rodda 1998). To increase a population one million-fold (from 2 to 2 million) requires about 20 doublings; spaced evenly over about 35 yr, this implies a doubling time of about 1.75 yr, or an annual population increase of about 40% \( (r = 0.396) \). This value reflects a population expanding at an extraordinarily high rate (and the population must have maintained this average rate of increase for more than 30 yr). It may be unrealistic to assume that such a high rate would prevail in new colonizations, both because conditions elsewhere may not be as optimal as in the Marianas (Rodda et al. 1999) and because countermeasures will undoubtedly be taken to suppress new Brown Tree Snake colonizations.

**RESPONSE TO MANAGEMENT**

Brown Tree Snake control has been practiced primarily to prevent dispersal of snakes from Guam to other islands (interdiction). A secondary effort has been directed at restoring wildlife populations on Guam through landscape-level control of snake populations (wildlife restoration).

The primary effort (interdiction) has been focused on ports and airports (Rodda et al. 1998); it has been very successful in those areas (Vice and Pitzler 2002). The primary tools for interdiction control currently in use include visual searches, dog-aided searches, and snake traps (Rodda and Fritts 1992; Engeman and Linnell 1998, Engeman et al. 1998, Rodda et al. 1998, Campbell et al. 1999, Vice and Engeman 2000, Vice and D. S. Vice, 2004, Vice et al. 2005). Prey base reduction and snake barriers have a secondary role in current interdiction management (references just cited, Perry et al. 1998). One employee can maintain about 175 traps or one search dog. Search dog teams (one handler plus one dog) interdict 60–90% of snakes planted in cargo that a team inspects (Engeman et al. 2002; J. Gibbons, M. A. Hall, D. S. Vice, and C. S. Clark, pers. comm.). Permanent removal of Brown Tree Snakes is believed to require barriers to prevent snake recolonization (Savarie et al. 2001); unfortunately, barriers that can withstand the severe hurricanes to which Guam is routinely subjected are expensive ($400+/m), and implementation has been correspondingly limited.

Several toxicants are effective against Brown Tree Snakes (Savarie and Bruggers 1999), and acetaminophen is registered for this use both in aerial broadcast and bait stations (Savarie et al. 2001, Johnston et al. 2002, National Wildlife Research Center 2003). Bait station use is indicated for interdiction purposes or for wildlife restoration where vulnerable nontarget species are present and ground-based access to bait stations is practical. Aerial broadcast offers the prospect of landscape-level control, even in areas without easy access. Costs have not yet been published for aerial broadcast of acetaminophen, but bait station use appears roughly comparable in cost with that of snake traps (Daniel Vice, pers. comm.). Complete cost benefit analyses of available control tools have not yet been published. No Brown Tree Snake eradications have been completed on any scale, but may be possible, at least on a modest scale (Rodda et al. 1999, 2002).
Natural Enemies

One management tool that has not been developed for Brown Tree Snakes is natural enemies. The traditional form of control would be to obtain a natural predator from the snake’s native range. There are no known predators that specialize in consuming Brown Tree Snakes (Rodda et al. 1999c). There are two snakes that are known to consume Brown Tree Snakes opportunistically: Ophiophagus hannah and Stegonotus cucullatus. The former is commonly known as the King Cobra and would not be an appropriate control tool (for obvious reasons), whereas the latter is a harmless egg-eating snake. It coexists with the Brown Tree Snake in areas where the Brown Tree Snake is common, suggesting that whatever depressive effect it has is relatively modest.

Another approach would be to introduce a poisonous prey item for the Brown Tree Snake. An obvious candidate would be the poisonous toad Bufo marinus. This exotic prey already occurs on Guam without any apparent impact on the snake.

Several authors have explored the possibility of using a disease or parasite found in the snake’s native range (Whittier and O’Donoghue 1998, Telford 1999, Caudell 2001, Caudell et al. 2002, Jakes et al. 2003), with the uniform conclusion that all known natural diseases and parasites are not known to have demographic significance (or even clinical significance) as a biocontrol agent.

Finally, some have suggested that novel disease agents (those not found in the wild and therefore for which the Brown Tree Snake might not have evolved resistance) might be of some value (Dobson 1988, Holzmann 1999). A blue-ribbon panel of pest management experts recently explored this option (Colvin et al. 2005) and concluded that the risks of this approach and the expense inherent in safe development were beyond the reach of available funds and should not be pursued at this time.

Prognosis

The existing U.S.D.A. Wildlife Services program to interdict Brown Tree Snakes leaving Guam has been highly successful (Colvin et al. 2005). This has greatly reduced the risks to Guam’s trading partners. Concern surrounds rapidly escalating military activities on Guam (and concomitant increases in cargo movements from Guam), the traditional difficulty in sustaining successful governmental programs over the long run (as complacency develops and new priorities emerge), and the difficulty in detecting and eradicating new colonizations of Brown Tree Snakes when they do occur. The islands of greatest risk are those that receive the greatest amount of cargo through Guam and those that lack native snakes (prey that have coevolved with snakes are less likely to be vulnerable: Rodda et al. 1999b); these include Micronesia generally, but especially the Northern Mariana Islands, and Hawai’i. Palau may be less at risk than are other parts of Micronesia, because Palau has native snakes that consume birds. The other localities in the Caroline Islands are at great risk, because they receive much cargo through Guam and do not have well-developed wildlife management infrastructures to inspect cargo or detect incipient colonizations of Brown Tree Snakes. The Northern Mariana Islands have an excellent Brown Tree Snake interdiction program, which is especially desirable because they get virtually all of their cargo through Guam. Hawai’i gets only a small percentage of its cargo through Guam, but that percentage will increase as direct surface shipping from Guam to Hawai’i begins in 2006. In addition, the total amount of cargo passing through Hawai’i is large, so that even though the percentage coming through Guam ports is small, the risks are substantial and continuous. One special consideration with Hawai’i is that it serves as a transportation hub for the entire Pacific region; thus an infestation of Hawai’i could have severe regional consequences.

At the current time the risk of Brown Tree Snake introductions to the Pacific from the snake’s native range is considered to be relatively minor, for two reasons: (1) the density of Brown Tree Snakes is generally much lower in the snake’s native range than it is in Guam, and (2) little cargo destined for tropical and subtropical areas originates in
the snake’s native range (e.g., Port Moresby, Honiara, Darwin, Halmahera). That could change as Indonesia industrializes or food shipments increase from Australia’s tropical regions. Fortunately, a good model interdiction program currently in place in Guam may provide a useful template for Brown Tree Snakes in their native range, or for other invasive species of similar biology.

**Literature Cited**


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